

Adsorption of Chromium(VI) from Aqueous Solution by Glycine Modified Cross-linked Chitosan Resin

Asmaa S. Hamouda¹, Sayed A. Ahmed¹, Nahla M. Mohamed¹, Mostafa M.H. Khalil^{2*}

¹Faculty of Postgraduate Studies of Advanced Sciences (PSAS), Beni Suef University, Beni Suef, Egypt.

²Chemistry Department, Faculty of Science, Ain Shams University, Abbassia, 11566, Cairo, Egypt.

THE ADSORPTION of Cr(VI) onto glycine-modified crosslinked chitosan (GMCCR) resin has been investigated. Batch experiments were performed to examine kinetics, adsorption isotherm, pH effect, and thermodynamic parameters. The effect of pH for the adsorption of Cr(VI) was studied at range from 2 to 6 and the equilibrium was accomplished within 150 minutes and maximum removal was achieved under the optimum conditions at pH 3. The result obtained from equilibrium adsorption studies are fitted Langmuir and Freundlich adsorption models and the data was found that the equilibrium data agreed very well with the Langmuir model. The maximum uptake was found to be 1.5 mmol/g (calc 1.75 mmol/g) at 25°C. Thermodynamic parameters for the adsorption system were determined at 298 K, 308 K and 318 K ($\Delta H^\circ = 22.85 \text{ kJ}\cdot\text{mol}^{-1}$; $\Delta G^\circ = -33.17$ to $-36.93 \text{ kJ}\cdot\text{mol}^{-1}$ and $\Delta S^\circ = 188 \text{ J}\cdot\text{K}^{-1}\cdot\text{mol}^{-1}$). The positive values of ΔH° and ΔS° suggest an endothermic reaction and increase in randomness at the solid-liquid interface during the adsorption. The negative values of ΔG° indicating a spontaneous adsorption process. The kinetic process was described very well by a pseudo-second-order rate equation.

Keywords: Modified chitosan, Adsorption, Kinetics, Thermodynamics, Cr(VI).

Introduction

Chromium occurs mainly in the oxidation states trivalent Cr(III) and hexavalent in the environment Cr(VI) state. Whereas Cr(III) is essential in human nutrition (specially in glucose metabolism) as well as for plants and animals at trace concentrations, the hexavalent Cr(VI) has been considered more hazardous to public health due to its mutagenic and carcinogenic properties [1]. It is also moving readily through soils and aquatic environments and is a strong oxidizing agent capable of being absorbed through the skin [2]. A wide range of physical and chemical processes is available for the removal of Cr(VI) from drinking water, such as electrochemical precipitation, ultrafiltration, reverse osmosis and ion exchange [3-5]. Most of these methods suffer from high operational costs. Therefore, it is necessary to develop new treatment processes that are not only effective, but also feasible in terms of cost [6-8]. Adsorption

is one of the most economically favorable and a technically easy method [9].

Chitosan has been reported to have high potential for adsorption of chromium(VI) [10,11]. It is an amino-polysaccharide constituted of both acetylglucosamine and glycosamine moieties. Chitosan has been widely applied to the fields of pharmacy processing biotechnology, food and analytical chemistry. Amino group and hydroxyl group in chitosan exhibit good ability to chelate metal ions. On the other side, chitosan represents suitable materials for binding of metal oxo-anion species because of numerous functional groups (e.g., -OH and -NH₂) with their suitable H-bond acceptor and donor sites. Adsorption capacity of chitosan can be improved by chemical means such as addition of functional groups, crosslinking and by physical conditioning of the biopolymer as gel beads or fibers [12-14]. Several chemical changes have been applied to chitosan in order to enhance

*Corresponding author e-mail: khalil62@yahoo.com

DOI: 10.21608/ejchem.2018.2989.1250

©2017 National Information and Documentation Center (NIDOC)

its uptake of Cr (VI) from the solution [15-19]. Novel chitosan resins possessing chelating moieties have been developed by using cross-linked chitosan resin as a base material. The development of chelating resin is important from the viewpoint of the collection and separation of metal ions [20]. The cross-linked chitosan is found to be very stable and maintain their strength even in acidic and basic solutions. These characteristics are very important for an adsorbent so that it can be used in a lower pH environment [21].

In the present work, we prepared modified glutaraldehyde-crosslinked chitosan glycine-type and used it to adsorb Cr (VI) ions in a batch system. The effects of the process parameters such as pH, temperature on the removal were investigated. In order to have better understanding of the adsorption process, some isotherm, kinetic and thermodynamic models were employed.

Experimental

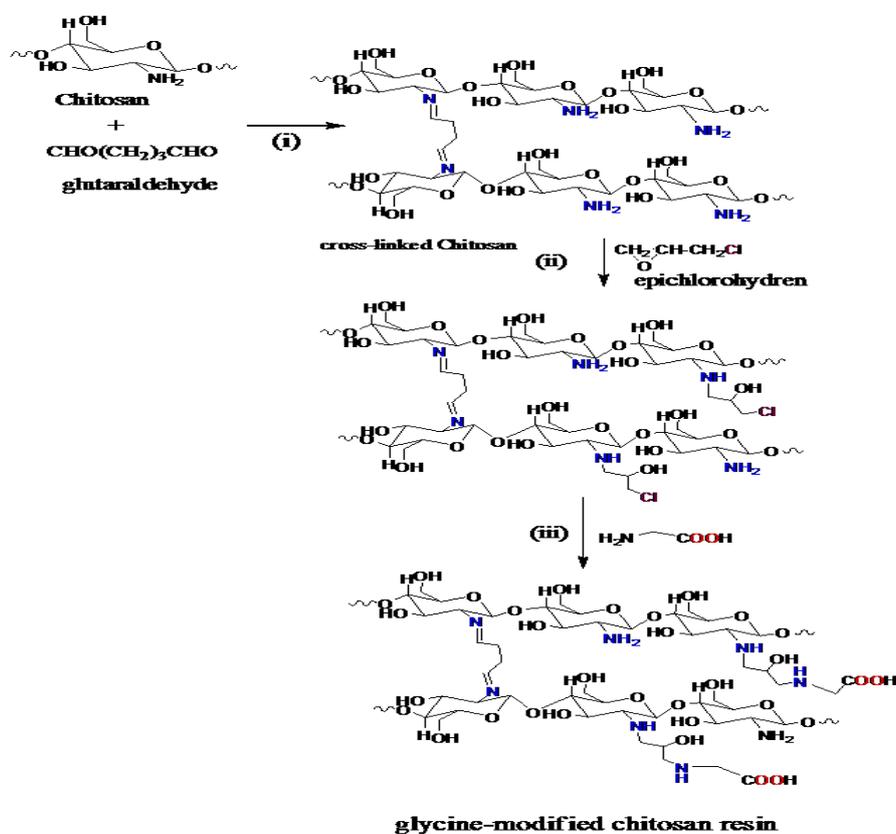
Chemicals

Chitosan (from crab shell), glycine, glutaraldehyde, isopropyl alcohol, epichlorohydrin, $K_2Cr_2O_7$ were Aldrich products. All other chemicals were Prolabo products and

were used as received.

Preparation of glycine modified chitosan resin

The glutaraldehyde-crosslinked chitosan glycine type was prepared as in literature [22]. Three grams of chitosan was dissolved in 20% aqueous solution of acetic acid and stirred until the solution became homogenous. Then 1 mL of glutaraldehyde solution (50%) was added and the solution was stirred with heating for two hours. The pH of the solution was raised to 6 and the obtained gel was washed with distilled water several times and kept to dry. The obtained cross-linked chitosan gel from the previous step was suspended in 60 mL isopropyl alcohol. Then 7 mL epichlorohydrin (62.5 mmol) dissolved in 100 mL acetone/water mixture (1:1 v/v) was added. The above mixture was stirred for 24 h at 60 °C. The obtained solid product was filtered off and washed several times with water followed by ethanol. The product and glycine (10 g) were suspended in dioxane (100 ml), then 40ml NaOH (1M) was added and the mixture was refluxed for 3 h. the final product was filtered and washed 3 times with ethanol and with deionized water. The synthesis steps are shown in Scheme 1.



Scheme 1

Characterization of the resin

Infrared spectra were performed using Nicolet 6700 FT-IR Spectrometer. The surface morphology of the adsorbent was visualized with a scanning electron microscope (JEOL-1200, Japan). The scanning electron microscopy (SEM) enabled us to direct observation of the changes in the surface microstructures of the adsorbent. X-ray diffraction (XRD, JCPDS No. 03-0921) was used to characterize the crystal structures of chitosan and GMCCR.

Water Regain

For water regain determination, resin sample was centrifuged for 30min at 1000rpm to remove excess water and then weighed. The sample was then dried at 50–60 °C until complete dryness then weighed again. To calculate this factor, the

$$W\% = \frac{(W_w - W_d) 100}{W_w} \quad (1)$$

Where W_w and W_d are weights (g) of the wet and dried resin, respectively. Water regain values are (37±3%). This value reflects the hydrophilic character of the resin type.

Uptake experiments using batch method

Preparation of solutions

Stock solution (1×10⁻² M) of chromium (VI) was prepared by dissolving 1.47g K₂Cr₂O₇ in 1L bi-distilled water. All batch experiments were carried out with adsorbent samples in a 250 mL conical flasks with 100 mL Cr (VI) aqueous solutions on a rotary shaker at 200 rpm. The concentration of Cr(VI) ions was determined spectrophotometrically at 540 nm using diphenylcarbazide as the complexing agent.

Effect of pH

The uptake of Cr (VI) by the investigated resins was studied at different pH values from 2 to 6. The pH was adjusted using HCl or NaOH. 0.1 g of investigated resin was placed in a series of flasks. To each flask 100 mL of Cr (VI) solution (5×10⁻³ M) was added. The contents of each flask were shaken for 150 min on a shaker at 200 rpm and at temperature 20 ± 1 °C at desired pH. The resin was separated from the solution by filtration. Then the residual concentration of Cr (VI) was determined.

Adsorption isotherms

Complete adsorption isotherms were carried out by placing 0.1 g portions of dried resin in a series of flasks containing 100 mL of Cr (VI) ions at pH 3. The temperature was thermostatically kept

at 25±1, 35±1 or 45±1°C and equilibrium time 150 min for studied resin. The residual concentration of Cr (VI) was determined. The adsorption data were treated according to Langmuir equation [23].

$$q_e = \frac{Q_{max} K_L C_e}{1 + K_L C_e} \quad (2)$$

Where q_e the adsorbed value of Cr (VI) ions at equilibrium concentration (mmol/g), Q_{max} is the maximum adsorption capacity (mmol/g) and K_L is the Langmuir binding constant which is related to the energy of adsorption (L/mmol), C_e is the equilibrium concentration of Cr (VI) in solution (mmol/L).

Its linearized equation is shown as below

$$\frac{C_e}{q_e} = \frac{C_e}{Q_{max}} + \frac{1}{K_L Q_{max}} \quad (3)$$

Plotting C_e/q_e against C_e gives a straight line with slope and intercept equal to $1/Q_{max}$ and $1/K_L Q_{max}$, respectively. The essential characteristics of the Langmuir isotherm can be expressed in terms of a dimensionless constant separation factor R_L that is given by

$$R_L = \frac{1}{1 + K_L C_i} \quad (4)$$

Where C_i (mmol/L) is the highest initial concentration of adsorbate and R_L values $0 < R_L < 1$ imply favorable adsorption. The Freundlich empirical relationship describes the multilayer adsorption of heterogeneous systems and assumes that different sites have several adsorption energies involved [18]. The linear model of the isotherm can be expressed logarithmically as

$$\log q_e = \log K_F + \frac{1}{n} \log C_e \quad (5)$$

Where, K and $1/n$ are Freundlich constants. The values of K and $1/n$, which roughly correspond to the adsorption capacity and the heterogeneity factor.

The thermodynamic parameters of adsorption reaction were obtained from the treatment of KL values at different temperature according to van't Hoff equation [24].

$$\ln K_L = \frac{-\Delta H^\circ}{RT} + \frac{\Delta S^\circ}{R} \quad (6)$$

where ΔH° and ΔS° are enthalpy and entropy changes, R is the universal gas constant (8.314

J/mol.K) and T is the absolute temperature (K). Plotting $\ln K_L$ against $1/T$ gives a straight line with slope and intercept equal to $-\Delta H^\circ/R$ and $\Delta S^\circ/R$, respectively. The Gibbs free energy of adsorption (ΔG°) at different temperatures was calculated using the following relation:

$$\Delta G^\circ = \Delta H^\circ - T\Delta S^\circ \quad (7)$$

Effect of contact time on the uptake

The effect of contacting time on the uptake of Cr (VI) by resins was carried out by placing 0.1 g of dry resin in a flask containing 100 mL of Cr (VI) solution at initial concentration of 5×10^{-3} M and pH 3. The contents of the flask were placed on a shaker at 200 rpm and at temperature 20 ± 1 °C. Five milliliters of the solution were taken at different time intervals and used to determine the residual concentration of Cr (VI). The adsorption data were treated according to the kinetic models,

Desorption experiments

For desorption studies, 0.15 g of GMCCR was loaded with metal ions using Cr (VI) solution (5×10^{-3} M) solution at 25 °C, pH 3 and contact time of 150 min, on a shaker at 200 rpm. After adsorption the GMCCR adsorbed with Cr (VI) ions were separated from the solution by filtration and then added into 30 mL 0.15N NaOH and stirred at 250 rpm for 30 min at 25 °C and the final Cr (VI) concentration was determined.

The adsorption–desorption cycles were

repeated consecutively five times to determine the reusability of sorbents. After each cycle of adsorption–desorption, sorbent was washed with distilled water and used in the succeeding cycle. The desorption ratio of Cr (VI) ions from GMCCR was calculated from the amount of Cr (VI) ions adsorbed on GMCCR and the final Cr (VI) ions concentration in the medium. Desorption ratio was calculated from the following equation:

$$\text{Desorption ratio} = \frac{\text{amount of Cr (VI) ions desorbed to the medium}}{\text{amount of Cr (VI) ions adsorbed onto the GMCCR}} \times 100 \quad (8)$$

Results and Discussions

FTIR and SEM analysis

To confirm the synthesis route of the glucine-chitosan crosslinked resin, FT-IR spectroscopy was used to determine the main characteristic functional groups on the adsorbent. The FTIR spectrum of chitosan powder is shown in Fig. 1(a). A broad peak centered at 3420 cm^{-1} was attributed to O-H and N-H stretching vibrations and the peak at 2882.9 cm^{-1} due to -CH stretching vibration in -CH, and -CH₂. The FTIR of chitosan showed bands that can be assigned as: 1650 cm^{-1} (-NH bending vibration in -NH₂), 1424 cm^{-1} (-NH deformation vibration in -NH₂), 1154 cm^{-1} (-CN stretching vibration), 1061.4 and 1031 cm^{-1} (-CO stretching vibration in -COH), and 897.6 cm^{-1} (-CN stretching vibration [25]. The FTIR spectrum of the (GMCCR) is presented in Fig. 1(b). After

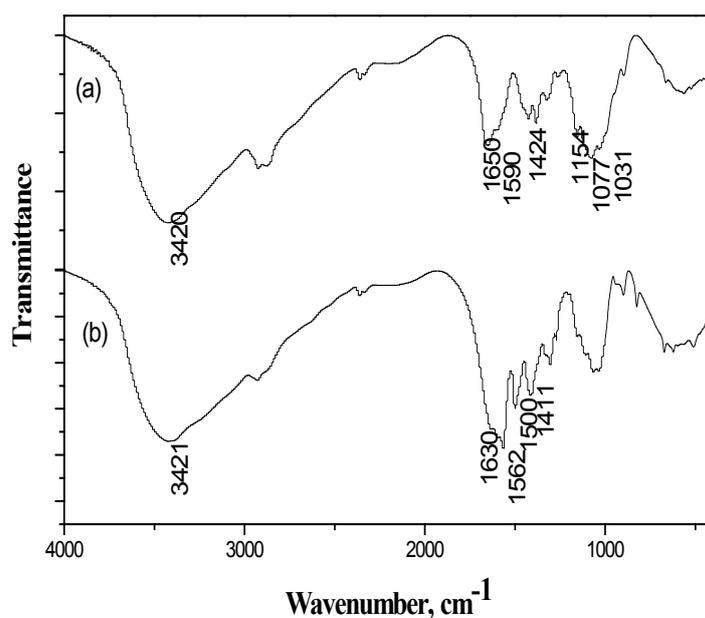


Fig. 1. FTIR of a) chitosan and b) glycine modified chosen resin(GMCCR).

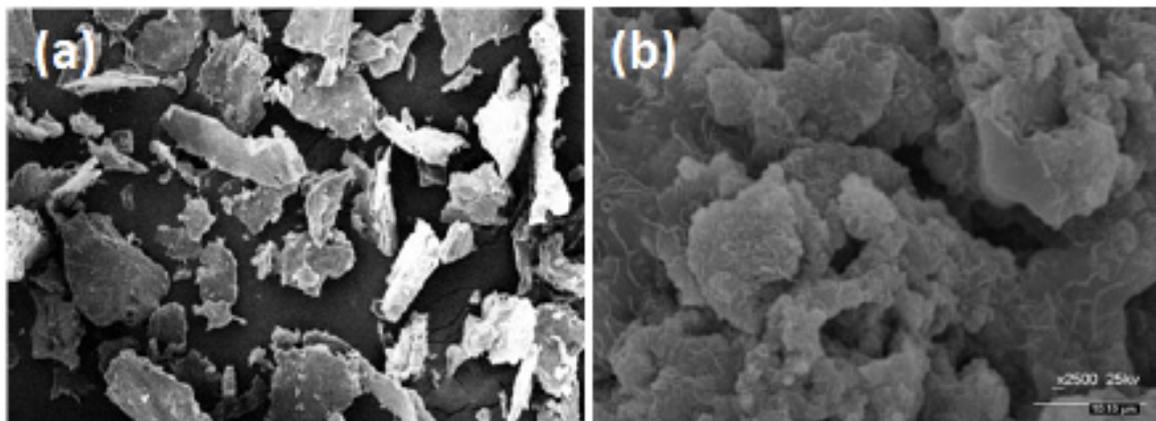


Fig. 2. SEM micrographs of a) chitosan and b) GMCCR resin (the surface morphology)

chitosan was crosslinked, an important peak in this region appeared at 1630 cm^{-1} ; it was attributed to the imine -C=N- band [26]. Meanwhile, The peak at 1562 cm^{-1} can be assigned to the C=N- stretching vibration, indicating that some of the -NH_2 groups were changed into the C=N- groups, and the new band at 1500 cm^{-1} and at 1424 cm^{-1} can be assigned to $\delta_s\text{NH}_3^+$ and COO- (stretching vibrations from glycine).

Figure 2 shows the SEM micrographs of the chitosan and GMCCR resin. It can be seen from Fig. 1(a) that the chitosan particles were mostly irregular in shape and had loose surfaces. After

they were crosslinked GMCCR becomes regular in shape and had dense surfaces and become rigid comparing to raw chitosan make it easier to separate from the solution (Fig. 1b). This result is in agreement with Huang et al [27].

Figure 3 shows the XRD patterns of chitosan and the GMCCR resin. In the chitosan XRD pattern, the peaks appeared at 11° and 20.1° correspond to a mixture of (1 0 1) and (0 0 2), and (0 0 1) and (1 0 0), respectively [28]. GMCCR exhibited the smaller characteristic peak of crystallinity than original chitosan, and it is clear that the crystallinity decreases after cross-linking using glutaraldehyde due to the Schiff's base

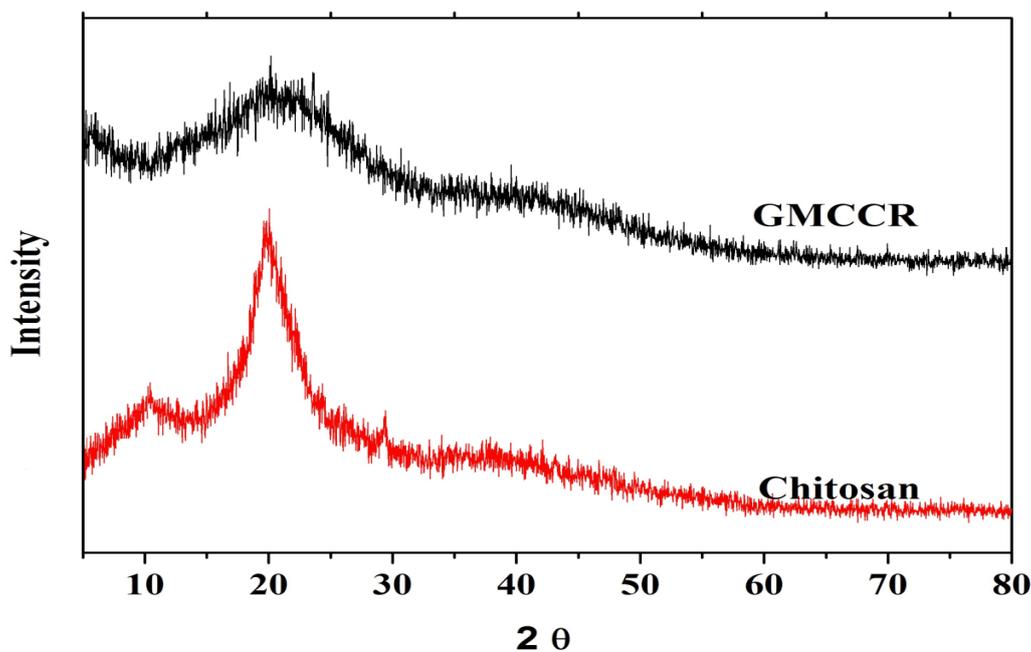


Fig. 3. The XRD spectrum of chitosan and GMCCR resin

TABLE 1. Langmuir constants for adsorption of Cr (VI) on Chitosan-Glycine resin.

Temp. (K)	Langmuir constants		Freundlich constants				
	$Q_{\max, \text{exp}}$ (mmol/g)	$Q_{\max, \text{calc}}$ (mmol/g)	K_L (L/mmol)	R^2	1/n	K_f	R^2
298	1.5	1.75	697	0.996	0.244	4.83	0.938
308	1.85	1.91	1025	0.997	0.215	5.17	0.996
318	2.10	2.31	1241	0.994	0.199	5.83	0.994

TABLE 2. Enthalpy, entropy and free energy changes for adsorption of Cr (VI) ions on Chitosan-Glycine resin.

$(H^\circ \text{ (kJ/mol}\Delta)$	$(S^\circ \text{ (J/mol.K}\Delta)$	Temperature, K		$\Delta G^\circ \text{ (kJ/mol)}$	$T\Delta S \text{ (kJ/mol)}$
22.85	188	298	-33.17	56.02	
		308	-35.05	57.90	
		318	-36.93	59.78	

formation between amino groups of chitosan and carbonyl groups of glutaraldehyde and subsequent reaction with glycine.

Effect of pH on the uptake

The pH is an important parameter for the adsorption process because of its influence on the adsorbent surface properties and ionic form of metal ion form in the solution. In acid medium, when the pH in the range of 2 to 6, HCO_4^- and $\text{Cr}_2\text{O}_7^{2-}$ are predominantly in equilibrium; in basic medium Cr(VI) exist in the CrO_4^{2-} form. The effect of pH on the uptake of Cr(VI) was studied at initial pH values of 2.0, 2.5, 3.0, 3.5, 4.0, 5 and 6. The modified chitosan showed higher uptake capacity towards Cr(VI) was achieved at pH 3, (Fig. 4). The important parameters for adsorption

of metal ions are the concentration of the counter ions on the functional group of the adsorbent and the degree of ionization of the adsorbate during the reaction. The decrease in the uptake at $\text{pH} < 3$ could be attributed to the higher concentration of Cl^- ions which compete the chromate anion. On the other hand, the decrease in the uptake value above pH 3 can be due to the lower extent of protonation of amino group with raising pH. The mechanism of interaction of chromate anion could be as anion exchange $\text{RNH}_2^+\text{Cl}^- + \text{HCrO}_4^- \leftrightarrow \text{RNH}_2^+ \text{HCrO}_4^- + \text{Cl}^-$ in the acid medium. Then, pH 3 was taken for the removal of Cr(VI) by GMCCR for further experiments. In alkaline conditions, it was found that the increasing OH^- anion brought a competitive sorption with chromate anions

TABLE 3. Kinetic Parameters for the adsorption of Cr (VI) on the studied Chitosan-Glycine resin.

Pseudo-First order				Pseudo-Second order			
k_1 (min^{-1})	$q_{e, \text{calc}}$ (mmol/g)	q_{exp} (mmol/g)	R^2	k_2 (g/mmol min)	$q_{e, \text{calc}}$ (mmol/g)	R^2	
0.0191	1.54	1.75	0.986	0.012	1.88	0.997	

TABLE 4. Adsorption–desorption cycles for GMCCR. (Adsorption conditions: volume, 100 mL; adsorbent dose, 0.15 g; initial concentration, $(5 \times 10^{-3} \text{ M})$; pH 3, contact time, 150 min; temperature, 25 °C; agitation speed, 200 rpm. Desorption conditions: 0.15 N NaOH; volume, 30 mL; contact time, 30 min; temperature, 25 °C; agitation speed, 250 rpm.).

Cycle number	adsorption	Desorption
1	100	92.5
2	92.3	90.9
3	89.8	85.1
4	86.6	78.3
5	82.5	76.4

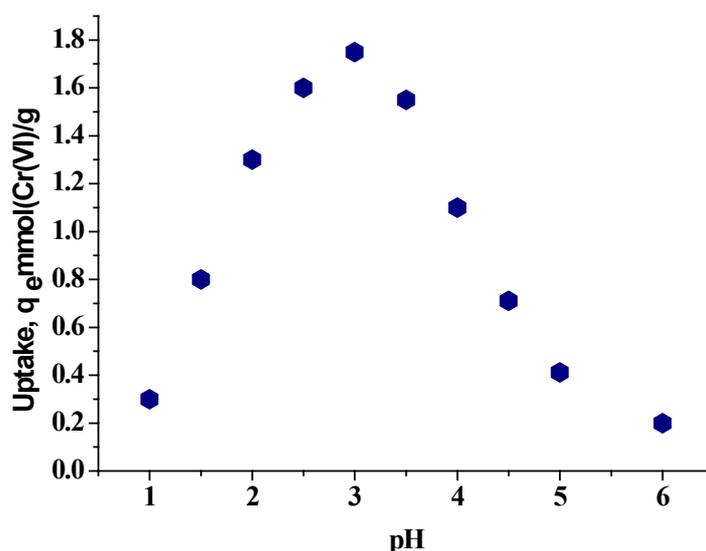


Fig. 4. Effect of pH on the uptake of Cr (VI) by Chitosan-Glycine resin at initial concentration of $5 \times 10^{-3} \text{ M}$ for; contact time 150 min at temperature 25°C.

leading to decrease in the sorption capacity of the modified chitosan towards Cr(VI) ions [29].

Adsorption isotherms

Adsorption isotherms are commonly used to reflect the performance of adsorbents in adsorption processes. Figures 5 and 6 represented the adsorption isotherms of Cr (VI) by (GMCCR) at pH 3 and at different temperatures using the Langmuir and Freundlich models, respectively. Both isotherms were fitted to experimental data, and the goodness of fit was compared. Isotherm

parameters for the Langmuir and Freundlich models for the(GMCCR) are reported in Table 1. At 25°C the maximum uptakes for (GMCCR) are 1.5mmol/g, . The values of K_L and Q_{max} for adsorption of Cr (VI) on resin was calculated from Langmuir isotherm (Fig. 5b). The values of Q_{max} are close to the experimental ones, and the values of R^2 reported in Table 1, which is a measure of the goodness-of-fit, confirm the better representation of the experimental data by Langmuir model. This indicates the homogeneity of active sites on the resin surface and suggests that the adsorption of

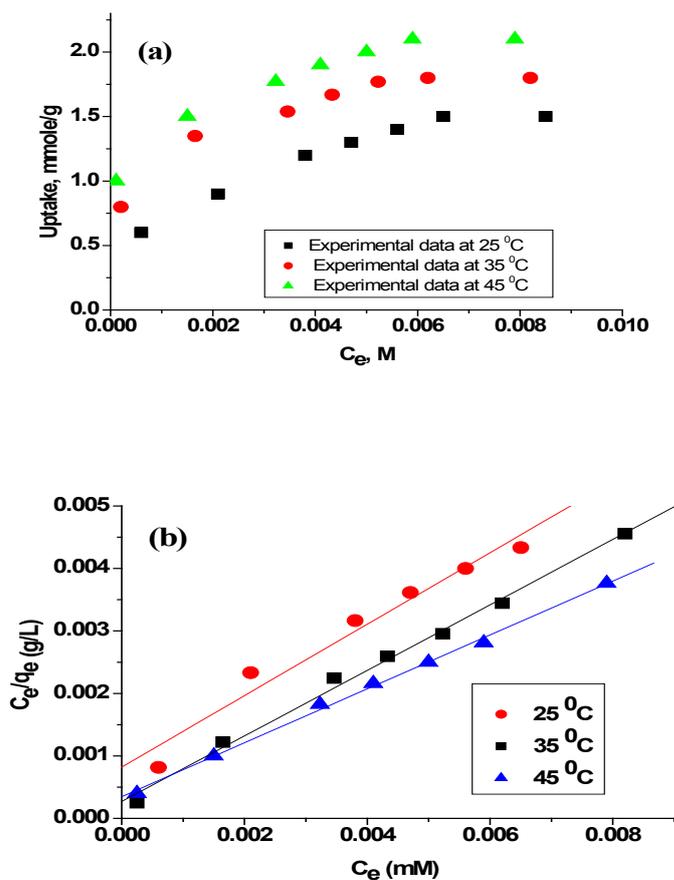


Fig. 5.a) Uptake Cr (VI) on (GMCCR) at different temperatures; pH 3; contact time 150 min, b) Langmuir isotherm.isotherm.

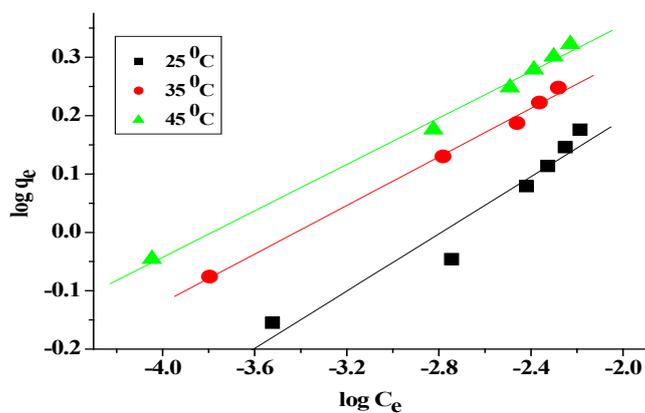


Fig. 6. Freundlich isotherms for the adsorption of Cr (VI) on (GMCCR) at different temperatures; pH 3; contact time 150 min.

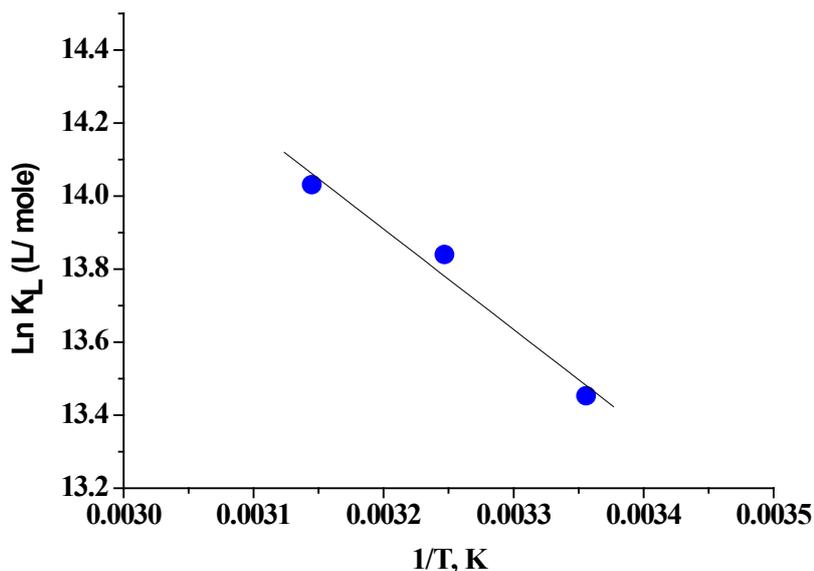


Fig. 7. Van't Hoff plots for adsorption of Cr (VI) on Chitosan-Glycine resin.

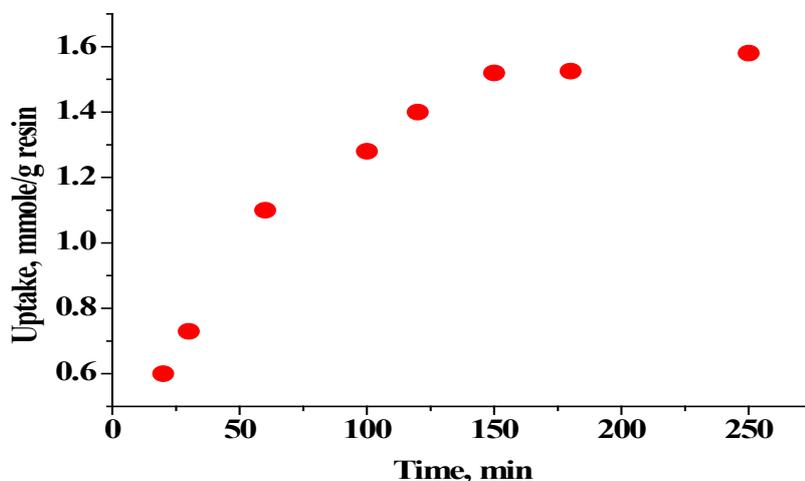


Fig. 8. Uptake of Cr (VI) by Chitosan-Glycine resin as a function of time at 25°C; pH 3; initial concentration of 5×10^{-3} M.

Cr(VI) on chitosan or GMCCR mainly occurred by monolayer reaction. The value of R_L in the present investigation was calculated to be 0.45 which indicates favorable adsorption of Cr(VI) on the resin.

Langmuir isotherm is found to be the applied isotherm for chitosan and modified chitosan by other authors [24, 30]. The observed increase

of Q_{max} and K_L with increasing temperature for (GMCCR) may be related to the increase of the stability of the complex formed between Cr (VI) and protonated amino groups (at pH 3) in the resin. The values of K_L gives a good idea about the ability of the resin for removing Cr (VI) from low concentrations. The reported values of K_L (Table 1) indicates the high removal efficiency of the resin for Cr (VI) from diluted solutions.

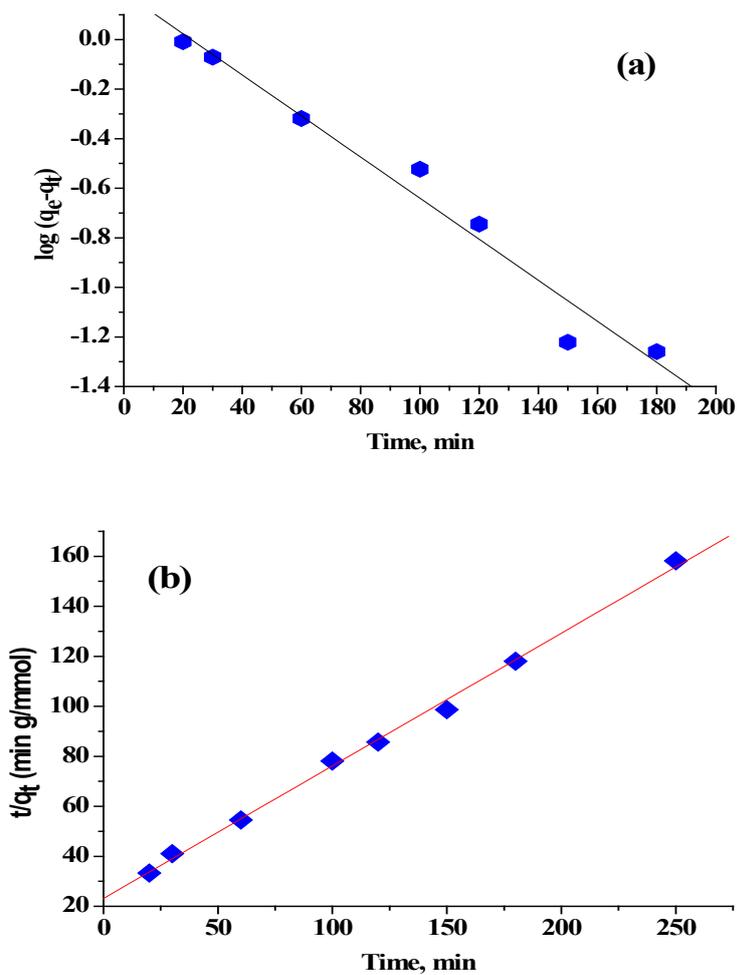


Fig. 9. (a) Pseudo first-orders and (b) Pseudo second-order kinetics of the uptake of Cr (VI) on (GMCCR) at 25°C; pH 3; initial concentration of 5×10^{-3} M.

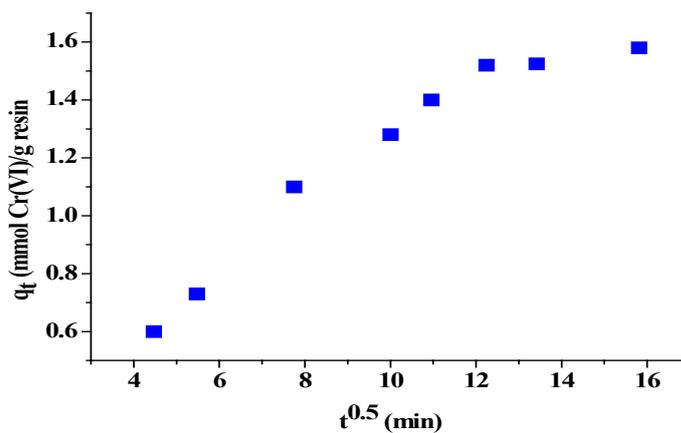


Fig. 10. The intraparticle diffusion kinetic model of the uptake of Cr(VI) by (GMCCR)

The reported values of ΔH° , ΔS° and ΔG° in Table 2 were calculated from Fig.7. The positive value of ΔH° indicates an endothermic adsorption process, and are coherent with chemical process, which confirms that the complex formation between Cr (VI) and amine sits on the resin surface. It is seen that the values of ΔG° decreases as the temperature increases. This implies that, the adsorption becomes more favorable at higher temperature [31]. The thermodynamic behavior may be attributed to the increased number of active sites as well as the endothermic nature of the reaction between resin active sites and Cr (VI) ions. This may also be reflected in the values of K_L . The values of K_L increase as the temperature increases, indicating higher affinity of the resin towards Cr (VI) at higher temperature.

Adsorption kinetics

The adsorption of Cr (VI) on (GMCCR) resin as a function of time at pH 3 and 25°C is shown in Fig. 8. Clearly, the equilibrium time was reached after approximately 150 min. The data in Fig. 8 were treated according to pseudo-first (Fig. 9a) and pseudo-second order kinetic models (Fig. 9b).

The pseudo-first order is models are expressed as [32]

$$q_t = q_e [1 - \exp^{-k_1 t}] \tag{9}$$

Its linearized equation is shown as below:

$$\log (q_e - q_t) = \log q_e - \left(\frac{k_1}{2.303}\right)t \tag{10}$$

where k_1 is the pseudo first-order rate constant (min^{-1}) of adsorption and q_e and q_t (mmol/g) are the amounts of Cr (VI) adsorbed at equilibrium and time t , respectively. The pseudo-second order model is expressed as [33]:

$$q_t = \frac{k_2 t}{1 + k_2 q_e t} \tag{11}$$

Its liberalized equation is shown as below:

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \left(\frac{1}{q_e}\right)t \tag{12}$$

where k_2 is the pseudo second order rate constant of adsorption ($\text{g mmol}^{-1} \text{min}^{-1}$). The kinetic parameters in both models are determined from the linear plots of $\log(q_e - q_t)$ vs t for pseudo first-order, (Fig. 9a) or (t/q_e) vs t for second order, (Fig.9b). The validity of each model is checked by the fitness of the straight line (R^2). The pseudo

first-order and pseudo second-order rate constants as well as equilibrium sorption capacities were reported in Table 3. According to the data shown in Table 3, for (GMCCR) resin, and the consistency of the between the calculated value and experimental value of q_e , pseudo second-order model is the more valid for the adsorption process than pseudo first-order one. This implies that the adsorption process proceeds according to pseudo second-order and depends upon both metal ion concentration and the textural properties of the resin. The intraparticle diffusion model indicated that the relationship between the concentration of Cr(VI) and the square root of time (Fickian diffusion law [24]; $q_t = K_t t^{0.5} + X$,) is not linear, Fig.8. This implies that the overall rate of adsorption for the resin is controlled by intraparticle diffusion with the boundary layer diffusion [34]. When intraparticle diffusion alone is the rate limiting step, then the plot of q_t versus $t_{0.5}$ passes the origin. When film diffusion is also taking place then the intercept is X , which gives an idea on the thickness of the boundary layer. From Fig. 8, it is evident that the adsorption process followed two steps. The first linear portion followed the boundary layer diffusion followed by another linear portion which represents the intraparticle diffusion [35].

Desorption studies

As shown in Table 4, the capacity of the GMCCR was an 7.7% decrease after the first cycle and a 11.2% decrease after the second cycle but it could still be maintained at 82.5% level at the fifth cycle. Complete desorption was not possible, perhaps due to the involvement of non-electrostatic forces between the GMCCR and the Cr (VI) ions. These results showed that the GMCCR can be successfully regenerated and repeatedly used in Cr (VI) ions adsorption studies without appreciable losses in their adsorption capacities.

Conclusion

The glycine-modified chitosan is found to have a maximum uptake of Cr(VI) of 1.5 mmol/g at 25 °C. Sorption of chromium on modified chitosan was influenced by the pH of the medium. Chromium sorption follows Langmuir isotherm. The nature of sorption process is spontaneous and endothermic. The mechanism of chromium sorption on the modified chitosan is governed by electrostatic adsorption and complexation. Regeneration of cross-linked magnetic GMCCR

obtained was achieved by using 0.15 M NaOH with efficiency of greater than 85%. The glycine modified chitosan is stable and good for chromium sorption and could be used for field applications.

References

1. Basha S., Murthy Z.V.P., Jha B.. Biosorption of hexavalent chromium by chemically modified seaweed, *Cystoseira indica*, *Chem. Eng. J.* **137**, 480–488 (2008).
2. Park S., Jung W.Y. Removal of chromium by activated carbon fibers plated with copper metal, *Carbon Sci.* **2**, 15–21 (2001).
3. Rengaraj S., Yeon K.H., Moon S.H. Removal of chromium from water and wastewater by ion exchange resins, *J. Hazard. Mater.* **87**, 273–287 (2001).
4. Benito Y., Ruiz M.L. Reverse osmosis applied to metal finishing wastewater, *Desalination*, **142**, 229–234 (2002).
5. Yurlova L., Kryvoruchko A. Removal of Ni(II) ions from wastewater by micellar-enhanced ultrafiltration, *Desalination*, **144**, 255–260 (2002).
6. Sharma C.D., Forster F.C. Removal of hexavalent chromium using sphagnum moss peat, *Wat. Res.* **27**, 1201 (1993).
7. Bailey S.E., Olin T.J., Bricka R.M., Adrian D.D. A review of potentially low-cost sorbents for heavy metals, *Wat. Res.* **33**, 2469 (1999).
8. Babel S., Kurniawan T.A. Low-cost adsorbents for heavy metals uptake from contaminated water: a review, *J. Hazard. Mater.* **B 97**, 219 (2003).
9. Karthikeyan T., Rajgopal S., Miranda L.R. Chromium (VI) adsorption from aqueous solution by Hevea Brasiliensis sawdust activated carbon, *J. Hazard. Mater.* **B124**, 192–199 (2005).
10. Modrzejewska, Z., Sujka, W., Dorabalska, M., Zarzycki, R. Adsorption of Cr(VI) on Cross-linked Chitosan Beads. *Sep. Sci. Technol.*, **41**, 111–122 (2006).
11. Zarzycki, R., Sujka, W., Dorabalska, M., Modrzejewska, Z. Adsorption of Cr(VI) on Chitosan Beads. *Chem. Inz. Ekol.* **9**, 1561–1570 (2002).
12. Alves N.M., Mano J.F. Chitosan derivatives obtained by chemical modifications for biomedical and environmental applications, *Int. J. Biol. Macromol.* **43**, 401–414 (2008).
13. Rojas G., Silva J., Flores J.A., Rodriguez A., Adsorption of chromium onto crosslinked chitosan, *Sep. Purif. Technol.* **44**, 31–36 (2005).
14. Khalil, M. M. H Atrees M. S., Abd El Fatah A. I. L., Salem H.& Roshdi R. Synthesis and application studies of chitosan acryloylthiourea derivative for the separation of rare earth elements, *Journal of Dispersion Science and Technology*, **39**(4), 605-613 (2018).
15. Rajiv Gandhi, M., Meenakshi, S. Preparation and characterization of La(III) encapsulated silica gel/chitosan composite and its metal uptake studies, *J. Hazard. Mater.*, **203–204**, 29–37 (2012).
16. Karthik, R., Meenakshi, S. Removal of hexavalent chromium ions from aqueous solution using chitosan/polypyrrole composite, *Desalin. Water Treat.*, **52**, 1–14 (2014).
17. Rajiv Gandhi M., Viswanathan, N., Meenakshi, S. Preparation and application of alumina/chitosan biocomposite, *Int. J. Biol. Macromol.*, **47**, 146-154 (2010).
18. Kousalya, G.N., Gandhi, M., Meenakshi, S. Sorption of chromium(VI) using modified forms of chitosan beads, *Int. J. Biol. Macromol.*, **47**, 308–315(2010).
19. Hua C., Zhang R., Bai F., Lu P., Liang X., Removal of chromium (VI) from aqueous solutions using quaternized chitosan microspheres, *Chinese J. Chem. Eng.* (2016), doi:10.1016/j.cjche.2016.08.024
20. Elwakeel K.Z., Removal of Cr(VI) from alkaline aqueous solutions using chemically modified magnetic chitosan resins, *Desalination*, **250**, 105-112 (2010).
21. Chiou M.S. and Li H.Y. Adsorption behavior of reactive dye in aqueous solution on chemical cross-linked chitosan beads, *Chemosphere*, **50**, 1095 (2003).
22. Khalil M.M.H., Al-Wakeel K. Z., Abd El Rehim S. S., Abd El Monem H. Efficient

- removal of ferric ions from aqueous medium by amine modified chitosan *J. Environm. Chem. Engineering* **1**, 566–573 (2013).
23. Langmuir, The adsorption of gases on plane surfaces of glass, mica and platinum, *J. Am. Chem. Soc.* **40**, 1361-1403 (1918).
24. Elwakeel K.Z., Abd El-Ghaffar M.A., El-Kosy S.M., El-Shorbagy H.G. Enhanced remediation of Reactive Black 5 from aqueous media using new chitosan ion exchangers, *J. Dispersion Sci. Technol.* **34**, 1008-1019 (2013).
25. Pentrák M., Pentrákova L., Radian A., Mishael Y. G., and Stucki J. W. Nitrate reduction by redox-modified smectites exchanged with chitosan. *Clays Clay Miner.* **62**, 403-414 (2014).
26. Li N., Bai R.A. Novel Amine-Shielded Surface Cross-Linking of Chitosan Hydrogel Beads for Enhanced Metal Adsorption Performance, *Ind. Eng. Chem. Res.* **44**, 6692-6700 (2005).
27. Huang R., Yang B., Liu Q. Removal of Chromium(VI) Ions from Aqueous Solutions with Protonated Crosslinked Chitosan, *J. Appl. Polym. Sci.* (2013) DOI: 10.1002/APP.38685.
28. Abou El-Reash, Y. G., Otto, M., Kenawy, I. M., & Ouf, A. M. Adsorption of Cr(VI) and As(V) ions by modified magnetic chitosan chelating resin. *International Journal of Biological Macromolecules*, **49**, 513–522 (2011).
29. Song W., Shi T., Yang D., Ye J., Zhou Y., Feng Y. Pretreatment effects on the sorption of Cr(VI) onto surfactant modified zeolite: mechanism analysis. *J Environ Manag.* **162**, 96–101 (2015).
30. Pang M., Kano N., Imaizumi H. Adsorption of Chromium (VI) from Aqueous Solution Using Zeolite/Chitosan Hybrid Composite, *J. Chem. Chem. Eng.* **9**, 433-441 (2015).
31. Jassal P.S., Raut V.P. and Anand N., Removal of Chromium (VI) ions from Aqueous solution onto Chitosan and Cross-linked Chitosan Beads, *Proc Indian Nat. Sci Acad.* **76**, 1-6 (2010).
32. Lagergren S., About the theory of so-called adsorption of soluble substances. *Kungliga Svenska Vet* **24**, 1-39 (1898).
33. Ho Y.S., McKay G., Pseudo-second order model for sorption processes, *Process Biochem.* **34**, 451-465 (1999).
34. Donia A.M., Atia A.A., Heniesh A.M., Efficient removal of Hg(II) using magnetic chelating resin derived from copolymerization of bithiourea/thiourea/ glutaraldehyde, *Sep. Purif. Technol.* **60**, 46–53 (2008).
35. Vasanth Kumar K., Ramamurthi V., Sivanesan S. Modeling the mechanism involved during the sorption of methylene blue onto fly ash. *J. Colloid Interface Sci.* **284**, 14-21 (2005).

(Received 26/2/2018;
accepted 10/7/2018)

امتزاز الكروم (VI) من المحلول المائي بالراتنج الكيتوزان المتقاطع المرتبط بالجليسين

أسماء سيد حمودة¹، سيد عبد القادر أحمد²، نهلة مخيمر محمد³، مصطفى محمد حسن خليل⁴

¹كلية الدراسات العليا للعلوم المتقدمة - جامعة بني سويف - بني سويف - مصر.

²قسم الكيمياء - كلية العلوم - جامعة عين شمس - القاهرة - مصر.

تمت دراسة امتزاز Cr (VI) على راتنج الكيتوزان المتشابك المعدل بالجليسين (GMCCR) و دراسة العوامل المؤثرة على هذا السلوك مثل: زمن الامتزاز- الأس الهيدروجيني للوسط- درجة الحرارة- تركيز المواد الممتزة- وحساب المعاملات الديناميكية الحرارية والكيناتيكية لتفاعلات الامتزاز. وجد أن الرقم الهيدروجيني الأمثل لامتصاص Cr (VI) يتراوح من ٢ إلى ٤ وتم تحقيق التوازن خلال ١٥٠ دقيقة وتم تحقيق أقصى إزالة تحت الظروف المثلى عند pH 3 النتائج التي تم الحصول عليها من دراسات امتزاز التوازن تم تطبيقها في نماذج امتزاز Langmuir و Freundlich ، وأظهرت النتائج أن بيانات الإتزان متفقة بشكل جيد مع نموذج Langmuir . تم الوصول الي الحد الأقصى لامتصاص ليكون ١,٥ ملمول / غرام (حسابيا ١,٧٥ ملي مول / غرام) عند ٢٥ درجة مئوية. تم تحديد المعاملات الديناميكية الحرارية لنظام الامتزاز عند ٢٩٨ كلفن و ٣٠٨ كلفن و 318 كلفن لتكون $\Delta H = 22,85^\circ$ القيم الإيجابية لكل من ΔH و ΔS تشير إلى تفاعل ماص للحرارة ، وزيادة في العشوائية على السطح الصلب أثناء الامتزاز وانها تعتمد على التغيير في الإنتروبي أكثر منها على المحتوى الحراري وكانت قيم ΔG سالبة تشير إلى عملية امتزاز تلقائية. تم وصف العملية الحركية من خلال معادلة الدرجة الثانية الزائفة بشكل جيد للغاية.